Modeling of the Regeneration Processes in Diesel Particulate Filters

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Abstract New simulation algorithm of soot oxidation in diesel particulate filter (DPF) is proposed. Two-dimensional finite differences scheme is used for calculation of the heat transfer in DPF (macro-scale). One dimensional model for treatment of the soot oxidation process in a separate filter channel (micro-scale) is based on a new expression for the oxidation reaction rate that is a modification of the known Bissett’s equation. This expression allows consideration of the non-linear dependence of oxidation reaction rate on a reagent concentration. For cases of catalytic coating a modification of the known two-layer model is suggested. A fraction of soot that has been oxidized with the help of the catalyst is not invariable in the first layer, but it exponentially depends on the distance from a channel wall of the particle trap. Soot oxidation by O2 and/or NO2 in diesel particulate filters is investigated for various cases of non-catalytic and catalytic oxidation. The filter clogging propagation from its periphery to the center at critical temperature is described theoretically at first. A comparison between various types of DPF regeneration methods is performed by using the proposed simple simulation algorithm and appropriate computer code.

Keywords Diesel particulate filter, Filter regeneration, Soot oxidation

1. Introduction

Diesel particulate filters are the most effective option to control the particulate matter (PM) emissions[1-4]. Today there are several types of DPF available on the market and differing mainly by the method of trapped soot removal usually called filter regeneration. It is convenient to classify regeneration methods as active (applying external or engine measures of gas temperature control at DPF inlet) or passive (usually employing catalytic means)[2],[5],[6]. Combinations of measures are also common and catalytic means can be applied in active regeneration systems. Between active regeneration methods utilizing engine measures of gas temperature control at DPF inlet the following means can be noted[2]: exhaust gas recirculation (EGR), injection timing retard, fuel post-injection, decrease of boost pressure, intercooler bypass. The external means of active regeneration include: fuel burners with full or partial flow, filter electric heaters, microwave heaters, injection of combustibles, catalytic or reactive species in the exhaust[7], electrochemical filter reactor[8], etc. Increase of the exhaust gas temperature to about 600°C (temperature of non-catalytic soot oxidation) by O2 presence, electric heating, etc. incur additional energy costs, require complex means of control and can createsignificant thermal stresses on filter components. Passive regeneration methods include: catalytic filter coatings[9], fuel-borne catalysts (FBC)[10], reactive species generation (e.g. NO2)[11]. The development of catalyzed DPF (CDPF) aims at attaining[2]: soot oxidation activity under moderate exhaust temperatures by exploiting direct (through oxygen transfer) as well as indirect (through NO2 generation) soot oxidation; reduced soot ignition temperatures compared to filter (CRT) or Catalyzed Continuously Regenerating Trap (CCRT), respectively)[3],[11]. CRT and CCRT systems usually use a diesel oxidation catalyst (DOC) located upstream a DPF to promote NO2 generation from the engine-out NO emissions. The mentioned above methods of DPF regeneration differ...
by their cost, reliability, minimal required exhaust gas temperature and NOx concentrations, etc.

Processes of DPF regeneration have been studied and described in numerous publications. The NOx carbon oxidation reaction follows the Langmuir-Hinshelwood mechanism and characterized by activation energy values far below those for C+O2 reactions and by non-linear dependence of the reaction rate, R, on NOx-concentration. Messerer et al.[12] has found experimentally that at NOx concentrations above 200 ppm the reaction rate increase with NOx fraction growth saturates. However, many modeling approaches apply the classic Bissett equation, and soot oxidation is assumed to be the first order reaction in NOx[13],[14].

For non-catalytic O2-carbon oxidation the activation energy$E_0$ value of 180kJ/mol was suggested by Konstandopoulos and Kostoglou[15]. Jacquot et al.[16] obtained $E_0/\Delta G^{\circ}$ = 45.5 kJ/mol for the reaction C+2NO2 $\rightarrow$ CO2+2NO and $E_0/\Delta G^{\circ}$ = 59.4 kJ/mol for that of C+NO2 $\rightarrow$ CO+NO. These values are adopted here. Some promising catalysts have been reported in[17],[18] which can catalyze the C+O2 reaction. Fino and Specchia[17] obtained activation energy of 123kJ/mol for catalytic O2-carbon oxidation by the development of a new soot combustion catalyst, PrCrO3 perovskite on CeO2. Palmisano et al.[18] produced nano-particles of ceria by combustion synthesis with catalytic activity of energy 111kJ/mol and temperature peak of carbon oxidation of 405 C. Wall-flow filters were studied by Konstandopoulos&Kostoglou[19] under conditions of O2 oxidation and by Konstandopoulos et al.[20],[21] with NOx- and O2-carbon oxidation. FBC that are supplied with the fuel were discussed in[21]. They were investigated, as well as wash-coated filters, by Hinot et al.[22]. It was found that Pt-doped soot and simultaneously filtered aerosols were both equally effective in reducing the oxidation temperature by up to 140-250°C. Conversely, the deposition of a thin soot layer of 5-10 micron thickness onto Pt produced only a slight temperature decrease of about 13-42°C. It should be noted that the FBC approach is inconvenient because of the inherent requirement of providing a continuous additive supply to the fuel, which also leads to its continuous consumption and to accumulation of metal oxide (ash) inside the DPF[23]. However, latest developments of FBC, e.g. Fe-based FBC described by Rocher et al.[10], aimed at mitigation of this drawback mainly by reduction of the required FBC dosing rates.

It is important to have an integrated simulation tool that will enable comparison between various types of regeneration methods, in order to select the optimal configuration for the given type of vehicle and its driving characteristics. Development of such a tool is the goal of the work presented in this paper. It is the same objective as in[24]. Nevertheless, authors believe that their proposed model is useful due to its simplicity, more accurate heat transfer prediction (because of using two-dimensional scheme) and obtained new results.

New expression for the oxidation reaction rate with non-linear dependence of it on a reagent concentration is proposed, which makes possible simulation of both O2- and NOx-carbon oxidation reactions. For cases of catalytic coating a modification of the known two-layer model is suggested. A fraction of soot that has been oxidized with the help of a catalyst is not invariable in the first layer, but it exponentially depends on the distance from a channel wall of the particle trap.

In macro-scale, for calculation of a temperature field in the whole filter the two dimensional finite differences scheme is used. A comparative study of various DPFRregeneration methods is performed using the developed model, corresponding computer code and based on the proposed new approach.

2. Simulation Model

2.1. Main Assumptions

The simulation model described here is for particulate filters with a structure of alternate plugged channels (blocked at their ends – wall-flow filters, proposed by Howitt&Montierth[25]). The continuum approach for this structure, which has been first introduced by Aris[26], is adopted for calculation of the filter temperature field, $T(z,x)$. Following Zhang et al.[27], a single temperature field (two-dimensional in the model) is used for the multiphase continuum consisting of three phases: exhaust gas, porous walls and particulate matter, excluding cases when part of the filter is blocked by the accumulated deposits. In the latter case there are only two phases remain: the structure walls and the particulate matter. Accumulation of soot in the wall is not considered. The findings of Mulone et al.[28] show that “theoretical wall capacity of particle collecting is never fully exploited allowing for a low impact of the wall onto backpressure” The assumption of the pressure derivative steadiness along axis z is adopted. The latter allowed considering the momentum balance integrally.

Other assumptions, for the whole filter as a continuum, are as follows: the deposits distribution over the filter length ($z$-axis) is uniform; the porous channel walls and soot deposits on them are taken into account by effective values of densities, radial thermal conductivities and volumetric specific powers; the continuum velocity is directed along the filter axis. This assumption is based on the fact that the exhaust gases from the inlet channels of the trap arrive to the outlet channels from all directions. Thus, there is no radial flow in the macro scale.

Obtained $T(z,x)$ is used for the soot deposit calculations in the channels along the radial coordinate $x$ after $T(z,x)$ averaging over axis coordinate $z$. These calculations for each radial coordinate $x$ are performed by applying the one-dimensional one- or two-layer model. One-layer model is used for investigation of the non-catalytic and FBC catalyzed carbon oxidation by nitrogen dioxide or/and oxygen. The investigation of filters that are coated with a
catalyst is performed by means of the two-layer mathematical model, similar to that developed by Konstandopoulos and Kostoglou[19] and differs from the latter by exponential dependence of the catalytic fraction of soot oxidation in the first layer on the distance from a channel wall of the particle trap coated by a catalyst. This model allows considering catalytic and non-catalytic oxidation by \( \text{O}_2 \) and by \( \text{NO}_2 \) separately or as a combination.

The following assumptions similar to those of Mohammed et al.[29], for the single channel, are made: the temperatures of the gas, the species concentration, the gas velocities and properties are taken as average across the channelcross-section; there are no variations of the species concentration in the exhaust gas along the length of the inlet channel; the gas is considered as ideal.

Based on these assumptions, the equations of: reaction rate; soot & ash deposit evolution; formation of \( \text{NO}_2 \)-carbon oxidation and \( \text{NO}_2 \)-carbon oxidation are competitive and according to Konstandopoulos & Kostoglou[15], the reaction rates for thermal and catalytic carbon oxidation are:

\[
R_{\text{O}_2} = \rho \cdot U \cdot w_X \cdot \left[1 - \exp\left(-S \cdot k \cdot \left(I - \phi_{\text{CO}}\right) \cdot d / U\right)\right] \tag{1}
\]

Here: \( \rho \)– gas density [kg/m\(^3\)]; \( U \)– gas velocity in the trap channels [m/s]; \( w_X \)– mass fraction of reagent \( X \); \( S \)– soot specific surface, equal to \( 5.5 \times 10^6 \) [m\(^2\) / m\(^3\)] according to[30]; \( d \)– soot layer thickness [m]; \( k \)– kinetic coefficient [m/s]; \( \phi_{\text{CO}} \)– non-catalytic carbon oxidizing coefficient [m/s] \( f_\text{act} \)– reaction order
\( \zeta \) is non-equal \( 1 \) – see eq. (1). In the case of carbon-by-\( \text{NO}_2 \)-carbon oxidation, the reaction rate is not really of the first order. Therefore, the reaction rate with \( f_\text{act} \) is dominant[20], it is possible to set the selectivity \( \zeta \) and an integration of the latter equation results in the form of eq. (1) for carbon-by-\( \text{NO}_2 \) oxidation in DPF for the reaction rate on the channel wall of the trap, as was obtained in[30]. However, the simulation results of the reaction rate with \( \zeta \) is 1 for carbon-by-\( \text{NO}_2 \) oxidation were found to be lower by two orders of magnitude than the experimental data of Messerer et al.[12]. This indicates that the reaction is not really of the first order. Therefore, the appropriate reaction is called in this paper as the “pseudo first order reaction”, and the value of \( \zeta \) adopted in this work is equal to 0.5 since it is best suited to the experimental results obtained in[12]. Therefore, the reaction rate equation for carbon-by-\( \text{NO}_2 \) oxidation can be written as:

\[
R_{\text{NO}_2} = \frac{\rho \cdot U \cdot \phi_{\text{NO}_2} \cdot \left[1 - \exp\left(-S \cdot k_{\text{NO}_2} \cdot d / U\right)\right]}{\sqrt{\nu_{\text{NO}_2}}} \tag{9}
\]

Here: \( \nu_{\text{NO}_2} \)is the \( \text{NO}_2 \) mole fraction. In fact, since the reaction \( 2\text{NO}_2 + C \Rightarrow 2\text{NO} + \text{CO}_2 \) is dominant[20], it is possible to set the selectivity \( \nu_{\text{CO}_2} \) as equal to 0.
The kinetic coefficient $k_{NOx}(used \ in \ eq. \ 9)$ and obtained by Jacquot et al.[16] is adopted here:

$$k_{NOx} = (k_{CO} \cdot p_{NOx}^\eta + k_{CO} \cdot p_{NO}^\lambda) \cdot L \cdot \lambda$$

(10)

where: $L$ – channel length[m]; $p_{NOx}$ is NO partial pressure[Pa]; $\eta = 1.13$, $\lambda = 1.05$ – parameters; and the kinetic coefficients $k_{CO}$ & $k_{NOx}$ are given by:

$$k_{CO} = 1.02 \cdot 10^{-3} \cdot \exp(-45,500 \ / \ RT)$$

(11a)

$$k_{CO} = 5.56 \cdot 10^{-3} \cdot \exp(-59,430 \ / \ RT)$$

(11b)

In recent works of Jeguirim et al.[32,33] the kinetic coefficients are obtained for soot oxidation reactions by $O_2$ and NO with and without coating by Ruthenium and Platinum on a channel wall of the trap. In the latter case oxidation starts from the conversion of NO into $O_2$. The following expression for the NO molar concentration $y_{NO}$ as a function of the NO molar concentration $y_{NOx}$ and the temperature, for the case of a catalytic oxidation with Pt, has been derived here from the work of Cooper and Thoss[34]:

$$y_{NO} = 0.00125 \cdot y_{NOx} \cdot T \cdot \exp[-\frac{T - 590}{135.33}]$$

(12)

The equation for soot deposition evolution in this case is obtained by adding on the right-hand-side (RHS) of eq. (8) the corresponding term:

$$\frac{d(\rho \cdot d)}{dt} = \frac{M_C}{M_{O_2}(1 - f_{CO}/2)} \cdot \frac{M_C}{\rho_{O_2}} \cdot \frac{M_C}{M_{NOx}(2 - g_{CO})} \cdot R_{NOx}$$

(13)

In the case of filters with catalytic coatings, the two-layer model is adopted as mentioned above. The thickness of the thin first layer, adjacent to the filter wall, can be taken as $d_1 = 10 \ \mu m$. The catalyst effect on the carbon oxidation occurs in the first layer only and is determined by a criterion for the fraction of soot that has been oxidized with the help of the catalyst: the index $\beta$. On the catalytic layer surface $\beta$ is equal to 0.5, as was suggested by Darcy et al.[35], and inside the second layer $\beta = 0$. In the first layer the following dependence of $\beta$ on a distance from the channel wall $x$ is proposed:

$$\beta(x) = \beta(0) \exp(-x / d_1)$$

(14)

where $\beta(0) = 0.5$.

If the soot deposit thickness is lower than or equal to $d_1$, all the equations for the one-layer model hold. For the one-layer model, non-catalytic or fuel borne catalyst oxidation, $\beta$ is equal to 0 or to a constant close to 1 (Konstandopoulos&Kostoglou[15]), respectively.

2.3 Modifications to the Basic Equations in the Two-Layer Model

The carbon oxidation process in the second layer is not affected by the presence of a catalyst in the case of catalytic coatings on the filter wall. Hence, equations (1 - 4, 8) for $O_2$ oxidation and (9 - 12) for NO oxidation are of the same form in this case. It is evident that in eqs. (1) and (9) for the corresponding reaction rates the soot layer thickness $d$ is replaced by $d_2 = d - d_1$. In eq. (8) for the soot deposit evolution, the term including $R_{O_2}^{soot}$ is set equal to 0 and the term

$$-\frac{M_C}{M_{NOx}(2 - g_{CO})} \cdot R_{NOx}$$

is added on the RHS (as for eq. 13) in the case of NO oxidation. In this case the soot evolution equation is:

$$\frac{d(\rho \cdot d_2)}{dt} = \frac{M_C}{M_{O_2}(1 - f_{CO}/2)} \cdot R_{O_2}^{soot}(d_2)$$

(15)

In the reaction rate equations (5, 6, 9) for the first layer, the consumption of the reagents $O_2$ or NO in the second layer must be taken into account. Thus, in these equations the corresponding reagent concentrations must be replaced by the following[29] for the $O_2$ oxidation process:

$$w_{O_2} = \frac{S \cdot k(1 - f_{CO}/2) \cdot d_2}{U}$$

(16)

and for the NO oxidation:

$$y_{NOx,1} = y_{NOx} \cdot \exp[-\frac{S \cdot k_{NOx}(2 - g_{CO}) \cdot d_2}{U}]$$

(17)

The values of $w_{O_2}$ and $y_{NOx,1}$ calculated by eqs. (16, 17) replace those of $w_{O_2}$ and $y_{NOx}$ in eqs. (5, 6, 9).

2.4 Other Equations

Pressure drop along the channel of the diesel particulate trap is expressed as in[36]:

$$\Delta p = \frac{\rho_0 \cdot Q}{2V} \cdot \left(b + D + 2 \cdot \frac{1}{2K_{soot}} \ln\left(\frac{b + D}{b - 2d_1}\right) + \frac{4P_2}{3} \cdot \left(\frac{1}{(b - 2d_1)^3} + \frac{1}{b^3}\right)\right)$$

(18)

where: $\mu$ - dynamic viscosity[kg/m·s]; $Q$ – exhaust gas volumetric flow rate[m³/s]; $V$ – filter volume[m³]; $b$– honeycomb filter cell width[m]; $K$, $K_{soot}$ – filter wall and soot permeability[m²]; $D$– channel wall and soot layer thickness, respectively[m]; $F = 28.5$ – constant.

The values of permeability, $K$ and $K_{soot}$ are taken from Versaevel et al.[37] and Huynh et al.[31], respectively: $K = 3.3 \cdot 10^{-13}$, $K_{soot} = 2.10^{-14}$[m²]. The non-homogeneity of the soot layer thickness in the radial direction was taken into account in the proposed model: the temperature field, $T(x,z)$, depends on the radial and longitudinal directions and the soot oxidation is a function of the $T$. The pressure drop along the whole trap was calculated by application of an electrical network analogy, in which $\Delta p$ was treated as a resistance. The values of the above-listed parameters are taken from[27] ($Q = 677$ [sem/h]; $b = 0.00015$ [m]; $D= 0.31$[mm]) or from specification of the considered vehicle. The radial velocity component $U_r(x)/$[m/s] in micro-scale is used for the soot deposit evolution calculations. It is calculated from the following expression:
inlet concentration increases by 50%. Equally, when the NO\textsubscript{2} feed decreases by 50%, the rate of soot oxidation is expected to decrease by about 20%. The adopted exponent value of $\zeta = 0.5$ (eq. 9) provided a good agreement of the predicted and previously published data on soot oxidation rate by NO\textsubscript{2}.

The model’s projection of the pressure drop across DPF was validated by the comparison of predicted data with those experimentally obtained by Mulone et al.[28] at the laboratory engine test bench for testing an 11 liter, 6-cylinder, inline heavy-duty diesel engine, equipped with a state of the art Johnson Matthey CCRT at steady-state conditions. This CCRT consisted of the DOC located upstream a catalyzed DPF. The main geometrical parameters of this aftertreatment system are shown in Table 2[28].

### Table 2. Main Geometrical Parameters of the CCRT Used for the Model Validation[28]

<table>
<thead>
<tr>
<th>Parameter</th>
<th>CCRT part</th>
</tr>
</thead>
<tbody>
<tr>
<td>Diameter [cm]</td>
<td>30.5</td>
</tr>
<tr>
<td>Length [cm]</td>
<td>13</td>
</tr>
<tr>
<td>Cell density [cell/cm\textsuperscript{3}]</td>
<td>62</td>
</tr>
<tr>
<td>Wall thickness [mm]</td>
<td>4</td>
</tr>
</tbody>
</table>

A comparison of the predicted data with those measured in[28] was carried out for the two steady-state regimes of CCRT operation: the soot loading (R10) and the regeneration mode (R100). R10 and R100 regimes corresponded to the 10 and 100% loads, respectively at the engine rating speed. Some relevant parameters of the experiment[28] that were used for the pressure drop predictions are shown in Table 3.

### Table 3. Relevant Parameters of the Experiment Applied in the Pressure Drop Predictions

<table>
<thead>
<tr>
<th>Parameter</th>
<th>R10 mode</th>
<th>R100 mode</th>
</tr>
</thead>
<tbody>
<tr>
<td>Fuel consumption [kg/h]</td>
<td>11.6</td>
<td>56.1</td>
</tr>
<tr>
<td>Intake air flow [sm\textsuperscript{3}/h]</td>
<td>436</td>
<td>999</td>
</tr>
<tr>
<td>CCRT inlet temperature [°C]</td>
<td>227</td>
<td>484</td>
</tr>
<tr>
<td>CCRT inlet NO [ppm]</td>
<td>124</td>
<td>361</td>
</tr>
<tr>
<td>CCRT inlet NO\textsubscript{2} [ppm]</td>
<td>40</td>
<td>117</td>
</tr>
<tr>
<td>CCRT inlet PM [mg/sm\textsuperscript{3}]</td>
<td>2</td>
<td>13</td>
</tr>
<tr>
<td>Soot packing density in the cake [kg/m\textsuperscript{3}]</td>
<td>40</td>
<td>40</td>
</tr>
</tbody>
</table>

A comparison of the measured and predicted pressure drop values for both soot loading and regeneration modes is presented in Fig.1. The obtained results show acceptable agreement between the calculated and experimental data. As can be seen, the model slightly underestimates pressure drop $\Delta p$ across CCRT. Maximal difference between the calculated and measured values of $\Delta p$ does not exceed 10%.

It is known that the overall pressure drop across a DPF is rather complicated and depends on losses through a porous wall and soot cake layer, losses to friction in the channel walls, inertial losses due to expansion/contraction and the Forchheimer effect[38]. Particles with different characteristics have various depths of penetration and packing densities, thus causing large variations in pressure.
drop due to wall loading and soot cake formation. Therefore, accurate prediction of the pressure drop across a DPF in real-world usage conditions is a very challengeable task. It requires very detailed information on engine operation parameters and exhaust emissions, which is hard to be achieved in vehicle fleet tests. Taking this into account, an attempt was made to roughly assess a pressure drop on a CRT over a long period in real-world usage conditions. Typical average values of engine parameters and exhaust emissions were used for this purpose.

Figure 1. Comparison of the model prediction with the experimental data of Mulone et al.[28] for the pressure drop across CCRT

A comparison was performed with the road test results of two CRTs installed on Mercedes Benz O-405 urban buses equipped by Euro-2 diesel engines (same model), Tartakovsky et al.[39]. During these tests the buses operated at their normal service with the real-world driving pattern typical for bus urban driving. The experiments performed in this study have proved the possibility of CRT successful operation in urban bus during about 2 – 2.5 years before the maintenance of ash clean-up. During the first 14 months of CRT operation on the bus, periodical measurements of pressure drop along the CRT were performed. These data have been compared with the results of the model prediction. The CRT that was studied here has the following filter parameters: channel width 2[mm], its length 305[mm], diameter 267[mm], wall thickness 0.43[mm]. It included the DOC and the downstream non-catalyzed DPF. The average value of the measured exhaust gas temperatures was 250℃.

Mass of the ash that comes from the lubricant oil and is accumulated in the DPF was taken into account. The mean values of the DOC inlet NO, NO₂ and PM emissions were 662 [g/h], 203 [g/h], and 220 [mg/sm³], respectively. In this work all the mass of PM was assumed to be only elementary or organic carbon.

Estimated values of the pressure drop across the CRT installed and operated in the urban bus, as described above, are shown in Fig. 2 and compared with the measured data obtained in the road tests[39].

The data shown in Fig. 2 demonstrate quite good agreement of the measured and predicted results. The maximal deviation between them does not exceed 20%, which is acceptable for this kind of comparison.

It can be concluded, based on the results of these comparisons, that the developed model can be used for assessment of the DPF suitability to the given vehicle type and its driving conditions.

Figure 2. Pressure drop on CRT – comparison of experimental and simulation data

4. Results and Discussion

Table 4. Main parameters used in the computations

<table>
<thead>
<tr>
<th>Parameter</th>
<th>Units</th>
<th>Value</th>
</tr>
</thead>
<tbody>
<tr>
<td>Gas flow rate</td>
<td>sm³/sec</td>
<td>0.188</td>
</tr>
<tr>
<td>Soot mass flow</td>
<td>mg/sec</td>
<td>4.2</td>
</tr>
<tr>
<td>Soot Packing Density in the Cake</td>
<td>kg/m³</td>
<td>40</td>
</tr>
<tr>
<td>Filtering effectiveness of DPF</td>
<td></td>
<td>0.99</td>
</tr>
<tr>
<td>Oxygen molar fraction in the exhaust gas</td>
<td>%</td>
<td>12</td>
</tr>
<tr>
<td>Filter channel width</td>
<td>mm</td>
<td>1.5</td>
</tr>
<tr>
<td>Filter channel length</td>
<td>mm</td>
<td>305</td>
</tr>
<tr>
<td>Wall thickness</td>
<td>mm</td>
<td>0.31</td>
</tr>
<tr>
<td>Filter diameter</td>
<td>mm</td>
<td>229</td>
</tr>
<tr>
<td>Channel wall permeability</td>
<td>m²</td>
<td>2E-13</td>
</tr>
</tbody>
</table>

The goal of the simulation model described in Section 2 and most of obtained results are the same as in[24]. Nevertheless, we presume that the proposed model is useful due to its simplicity and new effect observed because of the using the two-dimensional scheme for heat transfer calculation. The model was applied for study of the regeneration processes and performance of several DPFs. The results are presented below, for both catalytic and non-catalytic carbon oxidation. The process of oxidation by oxygen is taken into account also in the cases where the leading oxidation mechanism is by NO₂. In all the cases of DPF regeneration discussed here the parameters were considered as independent of time. The deposit of ash was
not taken into account. Main parameters of the DPF geometry and inlet flow that were used in the calculations described below are taken from[27],[40] and listed in Table 4.

Soot thickness parameter was chosen to allow study of the two-dimensional soot accumulation inside a DPF – see Fig. 3-8. This parameter is used for the similar purposes by Mulone et al.[28]. Sometimes the soot cake thickness is calculated to obtain an overall filter permeability, Rakovec et al.[38]. The developed model allows also prediction of the pressure drop across a DPF – Fig.1, 2. The latter parameter is widely accepted in industry to assess DPF behavior at various usage conditions[1],[41],[42].

4.1. Catalytic and Non-Catalytic O2-Oxidation

Fig. 3 shows the soot layer thickness \( d \) in the filter channels as a function of the temperature at equilibrium conditions for non-catalytic O2-oxidation and for two values of the activation energy. As can be seen, the levels of \( d \) under these conditions are reasonable and lead to pressure drop values on the DPF of less than 0.1bar at temperatures of 400\(^{\circ}\)C and 540\(^{\circ}\)C for \( E_{act} \) of 150 and 180kJ/mol, respectively. The latter value, suggested by Bissett[30], seems to be closer to the actual one: according to[12], non-catalytic O2-oxidation requires temperatures of about 600\(^{\circ}\)C.

![Figure 3. Soot layer thicknesses in the filter channels (in center and periphery) as a function of the temperature: non-catalytic O2-oxidation. Soot mass flow rate – 4.2 mg/sec](image)

4.2. Critical Temperatures Wherein Clogging of the Filter Periphery Starts

The curves in Figs. 3 and 4 illustrate thickness of the accumulated soot layer \( d \) for the filter center and filter periphery. The slight difference between them results only from heat losses in the filter periphery. The graphs for \( d \) as a function of the filter radius are nearly straight lines. For catalytic O2-oxidation (with catalytic coating) the behavior can be much different. Here the soot layer thickness, which is a function of the radial coordinate \( x \), and the resulting pressure drop are rather small, until the inlet gas temperature leads to the temperature at the filter periphery (which is lower than the temperature at the filter center due to heat losses on the outer filter surface) that is higher than a fixed critical value, \( T_{cr} \). At temperatures lower \( T_{cr} \), a clogging process will start from the filter periphery and continue to full clogging of the filter. The latter effect is a consequence of the filter conductivity increase in the clogged zone. When the soot layer thickness becomes higher than \( d_{c} \), the catalytic fraction \( \beta \) in the second layer becomes equal to zero, and the catalytic oxidation process is interrupted there. This normally leads to accelerated soot accumulation. If the filter wall temperature at the periphery is lower than \( T_{cr} \), and the temperature at the filter center is still above \( T_{cr} \), the soot accumulation at the filter periphery channels may drastically exceed that in the filter center. The calculations, using the constant \( \beta \) approach, show that for \( \beta = 0.5 \) (this value is taken according to[15]), \( d = 10 \) \( \mu \)m, (as in[19]) and \( E_{act} = 140\)kJ/mol (according to the review[37]) catalytic coating reduces \( E_{act} \) by 30-80kJ/mol and \( T_{cr} = 470^{\circ}\)C. In the case of using the variable \( \beta \) approach according to eq. (14), the same \( E_{act} \) and the same value of \( \beta(0) = 0.5 \), give \( T_{cr} = 511^{\circ}\)C. The former value is too ‘optimistic’: many published experimental results indicate that catalytic coating can reduce oxidation temperatures by 13 – 42\(^{\circ}\)C only, e.g.[22]. Indeed, for the non-catalytic O2-oxidation, as was mentioned earlier, the predicted value of acceptable temperature was 540\(^{\circ}\)C. The case of a more active catalytic coating with activation energy reduced down to 120kJ/mol (reduction of 60 kJ/mol compared to the value of 180kJ/mol that is relevant for the non-catalytic O2-oxidation) was studied too. For this case (the other parameters were kept as before) the value of \( T_{cr} = 400^{\circ}\)C was obtained. The most optimistic scenario of \( E_{act} \) reduction to 100 kJ/mol gives \( T_{cr} \) value of 303\(^{\circ}\)C.

![Figure 4. Soot layer thickness in the filter channels (in center and periphery) as a function of the temperature: O2-oxidation with a fuel-borne catalyst (\( \beta = 0.95 \)). Soot mass flow rate – 4.2 mg/sec](image)

Calculations of the filter clogging process near the critical temperature of 335 \(^{\circ}\)C are presented in Fig. 5 for the activation energy of 120kJ/mol (relevant for the catalytic soot oxidation). It should be noted here that in the real engine
operation an increase in the backpressure will inevitably lead to the raise of engine-out (DPF inlet) exhaust gas temperature. As a result, the process of soot accumulation in the vicinity of critical temperature will be appropriately slowed down. This phenomenon is not taken into account in the modeling and not reflected in Fig. 5.

4.3. FBC- and NO2 – Oxidation

Fig. 4 shows the soot layer thicknesses for carbon oxidation by means of FBC for two values of the activation energy: 100[40] and 123kJ/kmol. The latter has been obtained by Fino&Specchia[17] using perovskite on CeO2 catalyst. The results in Fig. 4 for the soot layer thickness d are obtained for β = 0.95. This value of catalytic fraction is chosen, as a conservative reproduction of the value β = 0.97 suggested by Konstandopoulos&Kostoglou[15]. As can be seen from Fig.4, acceptable temperatures for filter operation start at 220°C and 310°C, respectively. These temperatures are sufficiently low even for the latter case of $E_{act}$. The comparison of these results with the latest experimental data of Rocher et al.[10] on development of the Fe-based FBC shows that activation energy value of100kJ/kmol is too optimistic. The value of 123 kJ/kmol gives much more realistic results for the temperatures of soot burning with aid of nowadays available FBC.

The modelling results for non-catalytic NO2-oxidation process together with O2-oxidation are displayed in Fig. 6 for the two values of nitrogen oxide concentration: 100 and 200 ppm.

In this case, following eq. (12) the soot layer thickness dependence on temperature is non-monotonic. In these calculations the values of $E_{ONO2}^{(0)} = 45.5$ kJ/mol for $C+2NO_2\rightarrow CO_2+2NO$ and $E_{ONO2}^{(2)} = 59.4$ kJ/mol for the $C+NO_2\rightarrow CO+NO$ reaction[17] have been used. The corresponding soot depositions are acceptable (pressure drop values of less than 0.1 bar) at temperatures higher than 270°C and 325°C for NO concentrations of 200 and 100ppm, respectively. The former value of $T$ is very promising, whereas the latter is less acceptable, because lower temperatures prevail quite often in real driving patterns. Moreover, prediction results show that relatively low NO concentration (100ppm) yields too large soot layer thicknesses also in the medium $T$-range 450 – 510°C. In this range, the NO2-oxidation process is damped out, because of reduced NO to NO2 conversion (according to eq. (12)), while the O2-oxidation does not occur yet. At higher $T$, the two curves in Fig. 6 coincide, since the O2-oxidation process dominates. It is possible to completely avoid the disadvantage of increasing the soot layer thickness at medium temperatures of 450 – 510°C by means of a combination ofNO2- and O2- oxidation processes using a catalyst coating.

Figure 5. Filter clogging dependence on time. Soot mass flow rate – 4.2 mg/sec

Figure 6. Soot layer thickness in the filter channels as a function of the temperature: O2- and NO2-oxidation without catalyst. Soot mass flow rate – 4.2 mg/sec

Fig. 7a presents results for the catalytic coating, β(0) = 0.5, $d_1$ = 10 µm and $E_{act} = 140$ kJ/mol. An abrupt fall of both curves is observed when the temperature approaches the $T_{c_r}$= 511°C value of the catalytic O2 oxidation (the lower NO concentration the closer $T$ falls to $T_{c_r}$). At higher temperatures, the two curves in Fig. 7a merge, as also observed in Fig. 6, since the catalytic O2-oxidation dominates in this range. As can be seen from the comparison of data shown in Fig. 6 and 7a, the use of the catalytic coating that enables reduction of $E_{act}$ down to 140kJ/mol does not lead to any substantial improvement of soot oxidation at medium temperatures. Therefore, the possible impact of a more effective catalyst is studied. For example, the catalyst that enables reduction of the activation energy value down to123kJ/mol was considered. The predicted results of catalytic O2-andNO2-oxidation for the same values of β(0) = 0.5, $d_1$ = 10 µm and $E_{act} = 123$kJ/mol are shown in Fig. 7b.

As can be seen from Fig. 7b, this allows significant reduction of the soot layer thickness at medium temperatures and, therefore substantial improvement of the DPF regeneration performance. The results presented in Fig. 6, 7a,b are in a good agreement with those obtained by Soeger et al.[43] that at temperatures higher than 380°C there is an inflection point in the curve indicating a sharp increase in soot burning rate (in this work – sharp reduction of soot
deposit thickness – Fig. 7b). This behavior is explained by an increased contribution of oxygen-based soot combustion prevailing soot-by-NO₂ oxidation and compensates for the decreased NO₃ formation at temperatures higher than 300°C.

In addition to the discussed above study of catalytic O₂-andNO₂-oxidation, a similar process is investigated, where instead of catalytic coating a fuel-borne catalyst is used. Results obtained for the FBC case are presented in Fig. 8, for the same two concentrations of nitrogen oxide (100 and 200 ppm), as well as the same catalytic fraction values of β = 0.95 and activation energy $E_{\text{act}} = 120 \text{kJ/mol}$, that were used in the study of O₂-oxidation with FBC (see Fig. 4).

In the case presented in Fig. 8 (O₂-andNO₂-oxidation catalytic process with FBC), the minimal temperatures corresponding to acceptable pressure drop on the DPF are lower than those achieved with a non-catalytic or catalytic coating (coating) O₂-andNO₂-oxidation (see Figs. 6 and 7b, respectively), as well as with the O₂-oxidation by FBC for the same value of the activation energy $E_{\text{act}} = 120 \text{kJ/mol}$ (Fig. 4). The reason for this is the fact that in the case of FBC, the catalytic O₂-oxidation process starts at lower temperatures (compared to the catalytic coating) at the same activation energy of the reaction, since the fuel-borne catalyst affects the reaction in the whole soot layer. Therefore, in this case the acceptable minimal exhaust gas temperatures are 265°C and 285°C for NO concentrations of 200 and 100 ppm, respectively. The same results were obtained by Shejbal et al. [24]. In the latter, soot was significantly oxidized by O₂ and NO₂ at temperatures above 250°C. The disadvantages of FBC using were discussed above.

5. Conclusions

The developed model, based on the proposed new approach and corresponding computer code, enables comparative study of various DPF regeneration methods.

New expression for the reaction rate of soot oxidation with non-linear dependence of it on a reagent concentration is proposed. It makes possible simulation of both O₂- and NO₂-carbon oxidation reactions. For cases of catalytic coating a modification of the known two-layer model is suggested. According to this modified model, a fraction of soot that is oxidized with the help of catalytic coating is not invariable in the first layer, but it exponentially depends on the distance from a channel wall of the particle trap. Heat transfer in the particle trap as a continuum is considered by the two dimensional finite differences scheme.

Soot oxidation in diesel particulate filters by O₂, as well as NO₂ with O₂, have been investigated for various cases of non-catalytic oxidation and oxidation in presence of catalytic coating and fuel-borne catalyst. In the case of soot oxidation using catalytic coating, the phenomenon of critical temperature emergence was observed at first. It can be explained in the framework of the applied two-layer model by abrupt reduction of the reaction rate at the border of the second layer. Calculation results of soot oxidation by NO₂ with O₂ using catalytic coating show that soot is significantly oxidized at temperatures above 250°C. This finding is in a good agreement with data obtained by Shejbal et al. [24].

Dependence of the soot oxidation in presence of catalytic coating on activation energy is quite weak at relatively low temperatures of 250-350°C and strongly enhances with the temperature rise. This can be clearly illustrated for the case of NO concentration equal 200 ppm by the fact that the soot layer thickness of 160 micron is achieved almost at the same temperature of 250°C for both studied values of $E_{\text{act}}$ (140 and 123 kJ/mol). Conversely, at higher temperatures almost full soot burnout is achieved at 390°C for $E_{\text{act}} = 123 \text{kJ/mol}$ compared with 500°C for $E_{\text{act}} = 140 \text{kJ/mol}$.

The developed model allows selection of an acceptable soot oxidation method depending on available exhaust gas temperatures and nitrogen oxides concentrations. In future research the model can be further developed to cover other methods of soot regeneration that were not addressed in this study. The Bissett-type equations (such as eq. 1) can be a good approximation for modeling more complicated cases with more reactants.
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